# 1 Opportunities and Challenges for an Indonesian

# 2 Forest Monitoring Network

- 3 Francis Q. Brearley (f.q.brearley@mmu.ac.uk)<sup>1\*</sup>,
- 4 Wahyu C. Adinugroho (wahyuk2001@yahoo.com)<sup>2</sup>,
- 5 Rodrigo Cámara-Leret (R.CamaraLeret@kew.org)<sup>3</sup>,
- 6 Haruni Krisnawati (h.krisnawati@yahoo.co.id)<sup>2</sup>,
- 7 Alicia Ledo (alicialedo@gmail.com)<sup>4</sup>,
- 8 Lan Qie (qie.lan@gmail.com)<sup>5,6,38</sup>,
- 9 Thomas E. L. Smith (T.E.L.Smith@lse.ac.uk)<sup>7</sup>,
- 10 Fitri Aini (F.Aini@cgiar.org)<sup>8</sup>,
- 11 Fabien Garnier (garnierfabien2015@gmail.com)<sup>9</sup>,
- 12 Nurul S. Lestari (nurulsilva@gmail.com)<sup>10</sup>,
- 13 Muhammad Mansur (mansurhalik@yahoo.com)<sup>11</sup>,
- 14 Agustinus Murdjoko (agustinus.murdjoko.papua@gmail.com)<sup>12</sup>,
- 15 Satria Oktarita ((satriaoktaritanugraha@gmail.com)<sup>8</sup>,
- 16 Emma Soraya (esoraya@ugm.ac.id)<sup>13</sup>,
- 17 Hesti Lestari Tata (hl.tata@gmail.com)<sup>2</sup>,
- 18 Tatang Tiryana (tangtir@gmail.com)<sup>14</sup>,
- 19 Liam A. Trethowan (liamtrethowan@yahoo.co.uk)<sup>1,3</sup>,
- 20 Charlotte E. Wheeler (c.wheeler@ed.ac.uk)<sup>15</sup>,
- 21 Muhammad Abdullah (abdullah.m@mail.unnes.ac.id)<sup>16</sup>,
- 22 Aswandi (andiasw@yahoo.com)<sup>17</sup>,
- 23 Benjamin J. W. Buckley (bjwbuckley@yahoo.co.uk)<sup>18</sup>,
- 24 Elena Cantarello (ecantarello@bournemouth.ac.uk)<sup>19</sup>,
- 25 Iswan Dunggio (is\_onex@yahoo.com)<sup>20</sup>,
- 26 Hendra Gunawan (hendragunawan1964@yahoo.com)<sup>2</sup>,

- 27 Charlie D. Heatubun (charlie\_deheatboen@yahoo.com)<sup>3,12,21</sup>,
- 28 Diah Irawati Dwi Arini (irawati.diah@gmail.com)<sup>22</sup>,
- 29 Istomo (istomo19@gmail.com)<sup>23</sup>,
- 30 Tajudin Edy Komar (raminpd426@yahoo.co.id)<sup>2</sup>,
- 31 Relawan Kuswandi (r\_kuswandi@yahoo.co.id)<sup>24</sup>,
- 32 Zaenal Mutaqien (zaenal.mutaqien@lipi.go.id)<sup>25</sup>,
- 33 Sunitha R. Pangala (s.pangala@lancaster.ac.uk)<sup>26,27</sup>,
- 34 Ramadhanil (pitopang\_64@yahoo.com)<sup>28</sup>,
- 35 Prayoto (mrpray2000@gmail.com)<sup>29,30</sup>,
- 36 Antun Puspanti (puspantia@gmail.com)<sup>31</sup>,
- 37 Muhammad A. Qirom (qirom\_ma@yahoo.co.id)<sup>32</sup>,
- 38 Andes H. Rozak (andes.hamuraby.rozak@lipi.go.id)<sup>25</sup>,
- 39 Asep Sadili (asep.sadili@gmail.com)<sup>11</sup>,
- 40 Ismayadi Samsoedin (isamsoedin@yahoo.com)<sup>2,33</sup>,
- 41 Endah Sulistyawati (endah@sith.itb.ac.id)<sup>34</sup>,
- 42 Siti Sundari (ndariekologi@yahoo.com)<sup>11</sup>,
- 43 Sutomo (sutomo.uwa@gmail.com)<sup>35</sup>,
- 44 Agustinus P. Tampubolon (agus\_tampu@yahoo.com)<sup>2</sup>
- 45 Campbell O. Webb (cowebb@alaska.edu))<sup>36,37</sup>
- 46 1 School of Science and the Environment, Manchester Metropolitan University, Chester Street, Manchester,
   47 M1 5GD, UK
- 48 2 Forest Research and Development Center, Research, Development and Innovation Agency of the
   49 Ministry of Environment and Forestry, Jalan Gunung Batu No. 5, Bogor 16610, West Java, Indonesia
- 50 3 Identification and Naming Department, Royal Botanic Gardens, Kew, Richmond, Surrey, TW9 3AE, UK
- Institute of Biological and Environmental Sciences, University of Aberdeen, St Machar Drive, Aberdeen,
   AB24 3UU, UK
- 53 5 School of Geography, University of Leeds, Leeds, LS2 9JT, UK
- 54 6 Department of Life Sciences, Imperial College London, Silwood Park Campus, Ascot, SL5 7PY, UK
- 55 7 Department of Geography and Environment, London School of Economics and Political Science,
   56 Houghton Street, London, WC2A 2AE, UK

57	8	Center for International Forestry Research, Jalan CIFOR, Situ Gede, Sindang Barang, Bogor 16114, West
58		Java, Indonesia
59	9	Sumatran Orangutan Society, Medan, North Sumatra, Indonesia
60	10	Center for Research and Development of Socio-Economic Policy and Climate Change, Research,
61		Development and Innovation Agency of the Ministry of Environment and Forestry, Jalan Gunung Batu
62		No. 5, Bogor 16610, West Java, Indonesia
63	11	Botany Division, Research Center for Biology, Indonesian Institute of Sciences, Cibinong Science Center,
64		Jalan Raya Jakarta–Bogor Km. 46, Cibinong 16911, West Java, Indonesia
65	12	Faculty of Forestry, Universitas Papua, Jalan Gunung Salju, Amban, Manokwari 98314, West Papua,
66		Indonesia
67	13	Faculty of Forestry, Universitas Gadjah Mada, Bulaksumur, Yogyakarta 55281, Indonesia
68	14	Department of Forest Management, Faculty of Forestry, Bogor Agricultural University, Kampus Institut
69		Pertanian Bogor, Darmaga, Bogor 16680, West Java, Indonesia
70	15	School of Geosciences, University of Edinburgh, Crew Building, The King's Buildings, Edinburgh, EH9
71		3JQ, UK
72	16	Department of Biology, Faculty of Mathematics and Natural Sciences, Universitas Negeri Semarang, Jalan
73		Sekaran–Gunungpati, Semarang 50229, Central Java, Indonesia
74	17	Forestry and Environment Research Development Institute of Aek Nauli, Jalan Raya Parapat Km. 10.5,
75		Sibaganding, Parapat 21174, North Sumatra, Indonesia
76	18	Borneo Nature Foundation, Jalan Bukit Raya Induk No. 82, Bukit Hindu, Palangka Raya 73112, Central
77		Kalimantan, Indonesia
78	19	Department of Life and Environmental Sciences, Bournemouth University, Poole, BH12 5BB, UK
79	20	Gorontalo Regency Research and Development Agency, Kompleks GOR David-Tony, Jalan Yusuf Hasiru,
80		Limboto 96211, Gorontalo, Indonesia
81	21	Research and Development Agency, Provincial Government of West Papua, Jalan Brigjen. Mar. (Purn.)
82		Abraham O. Atururi, Arfai, Manokwari 98316, West Papua, Indonesia
83	22	Manado Environment and Forestry Research and Development Institute, Jalan Raya Adipura, Kima Atas,
84		Manado 95259, North Sulawesi, Indonesia
85	23	Department of Silviculture, Faculty of Forestry, Bogor Agricultural University, Kampus Institut Pertanian
86		Bogor, Darmaga, Bogor 16680, West Java, Indonesia
87	24	Manokwari Environment and Forestry Research and Development Institute, Jalan Inamberi-Susweni,
88		Manokwari 98313, West Papua, Indonesia
89	25	Cibodas Botanic Gardens, Indonesian Institute of Sciences (LIPI), Jalan Kebun Raya Cibodas, Cianjur
90		43253, West Java, Indonesia
91	26	Lancaster Environment Centre, Lancaster University, Bailrigg, Lancaster LA1 4YQ, UK
92	27	School of Environment, Earth and Ecosystem Sciences, The Open University, Walton Hall, Milton Keynes
93		MK7 6AA, UK

- 94 28 Department of Biology, Faculty of Mathematics and Natural Sciences, Universitas Tadulako, Jalan
  95 Sukarno Hatta Km. 10, Tondo, Palu 94118, Central Sulawesi, Indonesia
- 96 29 Graduate School for International Development and Cooperation, Hiroshima University, 1-5-1
   97 Kagamiyama, Hiroshima 739–8529, Japan
- 98 30 Riau Provincial Environment and Forestry Office, Jalan Sudirman No. 468, Pekanbaru 28126, Riau,
  99 Indonesia
- 100 31 Research Institute for Natural Resource Conservation Technology, Research, Development and
   101 Innovation Agency of the Ministry of Environment and Forestry, Jalan Soekarno-Hatta Km. 38,
   102 Balikpapan 76112, East Kalimantan, Indonesia
- Banjarbaru Environment and Forestry Research and Development Institute, Research, Development and
   Innovation Agency of the Ministry of Environment and Forestry, Jalan Ahmad Yani Km. 28, 7 Landasan
   Ulin, Banjarbaru 70721, South Kalimantan, Indonesia
- 106 33 Belantara Foundation, Jalan Timor No. 6, Gondangdia, Menteng, Jakarta 10350, Indonesia
- 107 34 School of Life Sciences and Technology, Institut Teknologi Bandung, Jalan Ganesha No. 10, Bandung
  108 40132, West Java, Indonesia
- 109 35 Bali Botanic Gardens, Indonesian Institute of Sciences (LIPI), Candikuning, Baturiti, Tabanan, Bali 82191,
  110 Indonesia
- 111 36 Arnold Arboretum of Harvard University, 1300 Centre Street, Boston, MA 02130, USA
- 112 37 University of Alaska Museum of the North, 907 Yukon Drive, Fairbanks, AK 99775, USA
- 113 38 CURRENT ADDRESS: School of Life Sciences, University of Lincoln, Brayford Pool, Lincoln, LN6 7TS,
   114 UK
- 115 \* Correspondence: f.q.brearley@mmu.ac.uk
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# 131 Key Message:

132	Permanent Sampling Plots (PSPs) are a powerful and reliable methodology to help our							
133	understanding of the diversity and dynamics of tropical forests. Based on the current inventory of							
134	PSPs in Indonesia, there is high potential to establish a long-term collaborative forest monitoring							
135	network. Whilst there are challenges to initiating such a network there are also innumerable							
136	benefits to help us understand and better conserve these exceptionally diverse ecosystems.							
137	Keywords: tropical forests, carbon, data-sharing, dynamics, monitoring							
138	List of abbreviations: NFI = (Indonesian) National Forest Inventory, PSP = permanent sampling							
139	plot, REDD+ = Reducing Emissions from Deforestation and forest Degradation							
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### 142 1. Why monitoring tropical forests is important

143 Tropical forests are arguably the most important terrestrial ecosystems. Whilst occupying 144 around 15 % of the global land area, tropical forests store two-thirds of all the carbon in terrestrial 145 vegetation (Pan et al. 2013) and are the most important above-ground terrestrial carbon sink (Beer et 146 al. 2010; Pan et al. 2011; Soepadmo 1993). They house half the world's biodiversity and provide a 147 wide range of goods, including sources of new medicines, and ecosystem services including clean 148 and sustained water supplies, climate regulation and pollinators for crops (Cámara-Leret et al. 2016; 149 Ghazoul 2015; Peters et al. 1989; Ricketts et al. 2004). If suitably managed, tropical forests can 150 provide economic benefits through ecotourism, non-timber forest products, a sustainable source of 151 timber, and through carbon financing mechanisms for developing tropical countries such as REDD+. 152 Therefore, understanding where, how and why the world's tropical forests are changing is a key 153 question of global importance (Hansen et al. 2013; Pan et al. 2011).

154 The periods over which trees establish, grow and die (tens to hundreds of years) do not make for 155 rapid experimental tests of forest functioning. Instead, direct measurements of stands of trees over 156 long time periods are essential to truly understand forest processes and dynamics (Lutz 2015). 157 Permanent sample plots (PSPs) in which all trees are marked, identified and repeatedly measured 158 provide a series of direct observations on forest condition, dynamics and change over time. As 159 longitudinal data sets, PSPs offer an excellent opportunity to study forest dynamics, and to separate 160 short-term environmental impacts, such as drought, from long-term trends (Condit 1998). A forest 161 monitoring network is a series of PSPs using a consistent protocol - such networks allow an 162 assessment of numerous aspects of forest ecology, including biodiversity, biomass (analogous to 163 carbon stocks), regeneration, dynamics (including succession) and 'health'. Furthermore, forest 164 monitoring networks distributed along large geographical and environmental gradients allow 165 testing for the generality of factors controlling ecosystem functioning with increased statistical 166 power (Craine et al. 2007) and allow space-for-time analyses to project potential impacts of global 167 changes on forests.

Numerous high-impact studies based on PSPs as the fundamental measurement unit have greatly advanced our understanding of the function, biodiversity and evolution of tropical forests. For example, PSPs have provided clear evidence that the tropical forest above-ground carbon stock has been increasing over time (Lewis et al. 2009; Pan et al. 2011; Qie et al. 2017) but that the sink strength into this stock appears to be declining, at least in Amazonia (Breinen et al. 2015). The above studies were conducted in 'undisturbed', i.e. primary, forests but a major proportion of tropical forests have been disturbed by human activities. Fewer PSP networks have been established to study forest recovery from logging (Rutishauser et al. 2015; Sist et al. 2014) or from shifting cultivation (Chazdon et al. 2016) yet they are also providing valuable data. Furthermore, PSPs contribute vital datasets to improve our still poor understanding of patterns in tropical tree species richness (Slik et al. 2015; ter Steege et al. 2013), biogeography (Slik et al. 2018) and evolution (Baker et al. 2014) at multiple scales.
Field data collected on the ground from biogeographically well-replicated PSPs is also a prerequisite to calibrate remotely-sensed biomass mapping (e.g. Asner et al. 2010; Avitabile et al. 2016; Réjou-Méchain et al. 2014).

182 Permanent Sample Plots are a standard method but can be supplemented by biodiversity observing 183 networks such as the transect approach of the Asia-Pacific Biodiversity Observation Network 184 (Yahara et al. 2012, 2014). Larger PSPs (~50 ha), such as those established by the Centre for Tropical 185 Forest Science (CTFS, now ForestGEO), play an important role in furthering our understanding of 186 community ecological patterns as they monitor a larger number of smaller ( $\geq 1$  cm dbh) trees over 187 bigger areas. In contrast, smaller PSPs (usually 1 ha), such as those established by the Amazon 188 Forest Inventory Network (RAINFOR) and the Indonesian National Forest Inventory (see section 2) 189 offer extensive coverage that is more appropriate for a regional-scale forest monitoring network.

#### 190 2. Opportunities from permanent sample plots in Indonesia

191 Indonesia has the third largest area of tropical forest globally (following Brazil and D.R. Congo; 192 FAO 2015) including some of the largest extents of carbon-dense peat swamp forests. However, as 193 with other regions of the world, Indonesia's forests are undergoing rapid change and anthropogenic 194 disturbance (Abood et al. 2014; Gaveau et al. 2014) and around half the country's land area currently 195 supports primary forest (Kementerian Lingkungan Hidup dan Kehutanan 2015b; Margono et al. 196 2014). The forests of western Indonesia are highly productive and the dominant trees, the 197 dipterocarps (Brearley et al. 2016), have been favoured as commercial timber trees for many years 198 leading to the majority of accessible forests being brought into timber production. By contrast, the 199 forests of eastern Indonesia (especially Papua) contain few dipterocarps and remain more intact 200 owing to the rugged topography and isolation. More recent challenges include droughts and fires 201 associated with El Niño that have had marked impacts upon forest functioning (Page & Hooijer 202 2016; Slik 2004) and increasing forest fragmentation (Qie et al. 2017), yet large-scale analyses that test 203 for such impacts across Indonesian forests are largely absent.

204 Numerous PSPs have been established across Indonesia over the last c. 60 years but not all have been 205 maintained continuously. The earliest PSPs were established during the late Dutch colonial era, but 206 they were mostly in plantation forests to study tree growth and timber yield (Hart 1928; Von 207 Wulfing 1938). Among the first PSPs established in primary forest was the 1-ha plot set-up by 208 Willem Meijer (1959) to study the ecology of Gunung Gede's montane forests. Since then, PSPs 209 have played an important role in silvicultural research such as the STREK (Silvicultural Techniques 210 for the Regeneration of Logged-over Forest in East Kalimantan) project (Bertault & Kadir 1998). 211 The Indonesian National Forest Inventory (NFI) is a national program initiated by the Indonesia 212 Ministry of Forestry in 1989 (and implemented by the Directorate General of Forestry Planning) 213 utilizing PSPs. Through this program, PSPs were established systematically with a 20 x 20 km grid 214 across forested areas in Indonesia (< 1000 m above sea level) with the primary objective to monitor 215 the growth of timber stocks. In total, 2735 1-ha PSPs were established, although not all have been 216 monitored on more than one occasion (Kementerian Kehutanan 1996). Depending on the location, 217 the NFI plots were not necessarily located in logging concessions but all logging companies were 218 required to establish PSPs for monitoring growth and yield. In addition to monitoring timber 219 growth and yield, data from these PSPs has provided a basis for estimating carbon stocks and 220 changes associated with land-use change and forest management activities (Kementerian 221 Lingkungan Hidup dan Kehutanan 2016; Krisnawati et al. 2014, 2015).

222 Despite the large-scale coverage of Indonesia's NFI, the limited scientific access NFI offers to its data 223 and the few large-scale analyses that have resulted from the NFI's dataset limit our understanding of 224 the composition and functioning of Indonesia's tropical forests. Given the current threats to 225 Indonesia's forests, it is important that Indonesian and foreign scientists collaborate, with a 226 consolidated scientist-led forest monitoring network having the flexibility to address ecological 227 questions in a democratised and collaborative fashion, to jointly establish PSPs and analyse large 228 datasets spanning Indonesia's forests. To date, at least 150 ha of PSPs (besides those in the NFI) 229 have been established in primary forest, and are still maintained, in Indonesia (Table 1; Figures 1a & 230 2). Although these PSPs have different sizes, re-measurement intervals and measurement protocols 231 making direct comparisons challenging, they offer a starting point for developing an Indonesian 232 forest monitoring network with a standardised protocol. The density of sampling across the whole 233 of Indonesia is only about 3.4 ha of plots per 10<sup>6</sup> ha of primary forest and there are clear differences 234 in sampling density between different geographical regions (Table 1). The highest density (ratio of 235 plot area to primary forest area) of PSPs, by an order of magnitude, is found in Java and Bali (Table 236 1). Although the total area of PSPs is modest, the area of primary forest remaining is particularly 237 low on these islands leading to an overall very high sampling density. Of the outer islands, 238 Kalimantan has a high density of sampling – likely due to this being the centre of production forest 239 logging activity coupled with interest in its exceptional biodiversity since the times of early colonial 240 explorers. Sumatra has a similar sampling density and has also been heavily exploited for timber in 241 the past. Maluku also has a high sampling density but this is largely confined to Seram only. 242 Sulawesi and Nusa Tenggara have sampling densities comparable to the mean for the whole of 243 Indonesia (although note that there are only 2.5 ha of plots in Nusa Tenggara). Sampling density 244 for Papua is, by far, the lowest among the Indonesian islands; this is partly due to the large 245 remaining area of forest combined with difficulties in establishing PSPs in areas with challenging 246 access. Of these PSPs, nearly half have been measured on more than one occasion, thereby 247 markedly increasing their value for assessing forest functioning, with the median monitoring period 248 for those measured more than once being 8 years and the longest being 50 years (Fig. 2b). About 249 half of the plots that have been measured on more than one occasion are in Kalimantan (e.g. Qie et 250 al. 2017) so the total monitoring effort (plot area x monitoring length) at around 1300 ha years is an 251 order of magnitude greater than Java + Bali, Maluku, Sulawesi or Sumatra; none of the PSPs in Nusa 252 Tenggara or Papua have been re-measured (Fig. 2c). In addition, there are over 100 ha of PSPs in 253 disturbed forest (Fig. 1b); many of these are forests that have been logged; in this case, the 254 geographical foci are Kalimantan and Sumatra that have historically been important for timber and, 255 secondarily, in Papua where logging activities are currently expanding.

From the brief analysis above, it is clear that key geographical gaps exist mainly in eastern Indonesia particularly for Maluku (excepting Seram), Nusa Tenggara and Papua. In terms of climate, many areas of drier forest are under-represented (e.g. Timor), as is montane forest and forest over edaphic variants (such as *kerangas* or ultramafic geology). There are some PSPs found in peat swamp forests but many have been burnt or otherwise disturbed in recent years.

# 261 3. Challenges facing an Indonesian forest monitoring network

#### 262 3.1 Methods

Our aim here is not to provide a protocol or critique of methods for PSPs as this has been done
in previous work (Alder & Synott 1992; Burslem & Ledo 2015; Condit 1998; Ledo 2015; Phillips et al.
2016; Sheil 1995) but to note concerns with particular relevance to the Indonesian situation.

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*Plot size*: Too many PSPs reported in the Indonesian literature are simply too small to provide a generalisation of the area they study. Small plots (e.g. 0.04 ha) might be useful when installed in a series (e.g. 25) to provide data on forest biodiversity that does not require accurate scaling-up to larger areas. However, for a more in-depth assessment of forest biodiversity, the larger the area sampled, the greater the number of species captured due to a large number of rare species (Plotkin et al. 2000). Of the PSPs noted in our analysis, the median size is 0.25 ha whilst the most frequently sized plot is 1 ha (Figure 2a), which is comparable to forest monitoring networks on other continents 10

274 (Brienen et al. 2015; Lewis et al. 2009; Phillips et al. 2009, 2016). Small plots cannot accurately 275 predict forest biomass when scaled-up to a larger area due to a high edge:interior ratio that elevates 276 the relative importance of marginal boundary decisions (Burslem & Ledo 2015), a high coefficient of 277 variation between plots, and the likelihood they will not represent all forest stages (e.g. gap, building 278 and mature, sensu Whitmore 1998). Calibration of remote sensing data for large-scale forest 279 biomass mapping is more accurate if the PSPs can be ground-truthed accurately, which also requires 280 larger plots (Avitabile et al. 2016; Réjou-Méchain et al. 2015). Finally, small plots are also prone to 281 the 'majestic effect' where researchers may unconsciously select pristine forest with 'majestic' large 282 trees and avoid disturbed areas (Sheil 1995).

283 Frequency of measurement: Whilst the definition of a PSP is that trees will be re-measured at some 284 point in time, re-measurement intervals are not always regular. A typical re-measurement interval 285 is five years as this allows increases in tree size to be seen more easily. Whilst intervals of four to 286 ten years are appropriate for most recording purposes of PSPs (Sheil 1995), an increasing census 287 period leads to a greater likelihood of unobserved growth and therefore an underestimation of forest 288 productivity (Talbot et al. 2014). In cases of annual censuses, this will allow much better 289 predictions of forest dynamics in relation to annual climate fluctuations (Clark et al. 2010). 290 Dendrometer bands are a possible inexpensive alternative to increase measurement frequency 291 (Anemaet & Middleton 2013), but require much greater time investment at installation; such bands 292 can also avoid errors due to changes of the point of measurement. Of course, regularity of 293 re-measurement depends upon plot security and accessibility, and funding is a key determinant of 294 frequency of fieldwork activities (see section 3.3).

295 Parameters measured: Trunk diameter at breast height (usually 1.3 m) is the key parameter measured 296 as this can be incorporated into allometric equations to estimate tree and stand biomass (Chave et al. 297 2014); including tree height and crown size has been shown to increase accuracy of such equations 298 (Goodman et al. 2014). This is especially needed for dipterocarps that show different architectural 299 patterns compared to other tropical trees (i.e. taller for a given diameter: Banin et al. 2012). Forests 300 in Indonesia cover not only a wide range of soil and climatic types both within and across islands, 301 but also represent a great biogeographical range. Due to variable architectures that require local 302 height-diameter models for accurate biomass calculation, tree height data collected within plots are 303 extremely useful to improve biomass estimates (Ledo et al. 2016; Sullivan et al. 2018).

304 3.2 *Taxonomy* 

For assessment of species distributions and monitoring, accurate taxonomy, comparable among
 plots, is paramount. Good taxonomy is clearly challenging as PSPs often contain a large proportion

307 of sterile individuals. Indonesia is fortunate in having a large and well-maintained national 308 herbarium (Herbarium Bogoriense; BO) and a number of regional herbaria but many PSP 309 investigators do not routinely collect voucher specimens but rely on vernacular names instead. 310 Taxonomy takes on extra importance in a forest monitoring network where the aim is to make 311 comparisons among plots, but technological advances have a key role to play here (Baker et al. 2017; 312 Webb et al. 2010). While some Indonesian tree genera are reasonably well known, for example the 313 commercially important dipterocarps (Ashton 2004) many large genera such as Syzygium 314 (Myrtaceae) and Diospyros (Ebenaceae) have not been monographed. Similarly, digitization of 315 herbarium sheets at BO is ongoing but progress remains slow.

316 Vouchers for morphotypes can be made available across sites permitting analysis of distribution of 317 taxa without any formal species names, but obtaining the species name increases the value of the 318 voucher. Challenges for the taxonomy of PSP trees must be taken seriously, and we recommend the 319 following: i) make physical voucher collections of several specimens for each morphotype especially 320 where variation appears to be high and collect silica gel-dried samples for subsequent DNA 321 barcoding; ii) carry out routine visits to PSPs to collect fertile specimens as they become available; iii) 322 take high-quality photographs of the fresh vouchers (Webb et al. 2010) and share images and 323 metadata online; iv) cross-match vouchers and images across different sites to both validate formal 324 species name and provide distribution information; v) avoid the use of vernacular names, except as 325 an early step in the determination process yet value the experience of parataxonomists in the field 326 and technicians in herbaria; and vi) publish details of how taxon names were acquired, and give a 327 level of confidence in each formal name. Overall, it is far more useful to publish voucher collection 328 codes, images, morphotype codes and matches of morphotypes to images at other sites than to 329 simply list a botanical name with no additional information. Detailed primary data will also 330 greatly assist taxonomic specialists in the future as they work on the large, complex genera of 331 Indonesian trees.

#### 332 *3.3 Funding*

Funding presents a perennial challenge for forest ecological work, particularly in developing countries. Within Indonesia, PSP censuses are not considered as applied research, which receive priority for funding, although NFI plots have been allocated governmental funding. Current funding opportunities through the development of the Indonesian Science Fund (DIPI) and via the UK Newton Fund are positive in this regard. There is also the potential for knowledge-exchange partnerships with logging companies who may fund PSPs in their concessions although, as funders, they may consider themselves data owners (see section 3.4). REDD+ programmes bring similar opportunities for knowledge exchange and funding (Gibbs et al. 2007). Longer-term collaborations
between Indonesian researchers, companies and NGOs coupled with leading international expertise
are needed. Importantly, PSPs need to be locally owned, and international funding should be
invested for pump-priming and capacity-building in order to stimulate long-term funding input

344 from Indonesian sources into tropical forest monitoring.

# 345 3.4 Data-sharing

346 Developing an integrated picture on changes in forest functioning and biodiversity across a 347 forest monitoring network requires the willingness to share data among researchers. Nevertheless, 348 data-sharing can present various challenges. There are a number of data-sharing models in tropical 349 ecology, ranging from the informal to the formal with rigid data-sharing arrangements such as 350 ForestPlots (López-González et al. 2011). What is shared can vary from whole plot data to only the 351 numbers required for a particular analysis. Issues over intellectual property are of considerable 352 concern and unwillingness to share data is often linked to concerns about the loss of control over 353 such data and the lack of professional recognition or reward (Enke et al. 2012; Fecher et al. 2015). 354 Furthermore, clarifying who is the 'owner' of data is essential. In some cases, the funder (often a 355 logging company) may claim ownership, in others, such as the Indonesian NFI, public access to the 356 data is limited. Any forest monitoring network needs clear guidelines on the sharing, use and 357 publication of shared data and an obvious reward system for sharing (i.e. co-authorship).

358 Although in-country data owners will regularly be included as co-authors in large-scale data 359 analyses, the lead authors have almost always been researchers from extra-tropical countries. 360 Echoing the sentiments of Ruslandi et al. (2014), we note that simply 'out-sourcing' data analysis to 361 extra-tropical researchers is still far from the goal of building local research capacity. Lack of 362 institutional support and incentive may deter tropical scientists from becoming leading authors, but 363 this appears to be changing lately with Indonesian institutions increasingly rewarding staff 364 publishing in international journals. Investing in capacity-building and knowledge exchange to 365 support Indonesian scientists to take leadership roles in agenda setting is also important in the 366 medium term.

#### 367 3.5 Land tenure and community engagement

368 Once a series of PSPs has been established it is important to maintain a commitment to re-measure 369 plots and obtain funding to do so. However, the location and accessibility of plots needs to be 370 considered for long-term measurements. Ideally, plot locations should not be too remote to make

accessibility challenging and not too close to settlements put plots at risk from disturbances. If new

372 PSPs are installed, there should be secure land tenure (Soraya 2011) to offer protection from land-use 373 change and fire risk – particularly in peat swamp forests (Page & Hooijer 2016). Of the PSPs noted 374 (Table 1; Figures 1 & 2), less than half are within formally protected areas (e.g. National Parks or 375 Nature Reserves); of those that are not, the presence of researchers may help in protecting them to 376 some degree (Laurance 2013). In areas where forest land-use classifications may jeopardise studies, 377 it may be possible to re-designate land classifications (e.g. Kawasan Hutan Dengan Tujuan Khusus 378 or 'Special Use Forests'). Local stakeholder engagement is key, and local communities should be 379 considered as valuable collaborators who value the presence of PSPs and can be employed to collect 380 good quality data (Theilade et al. 2015). There are multiple opportunities for synergies between 381 local communities, logging companies and scientists, with NGOs often in a strong position to act as 382 facilitators. Still, unless direct payments to forest owners are established for missed opportunities 383 of economic development, communities may well continue to prefer the economic benefits offered 384 by logging companies over those from researchers or conservationists (Novotny 2010).

# 385 4. Translating results from PSPs to forest policy and conservation

Quantification and assessment of carbon stocks in forests underpins international policies to mitigate carbon dioxide emissions such as the REDD+ program (Gibbs et al. 2007) and the recommendations of the Intergovernmental Panel on Climate Change (Watson et al. 2000). For example, Indonesia's forest reference emission level submitted to United Nations Framework Convention on Climate Change (Kementerian Lingkungan Hidup dan Kehutanan 2015a, refined in 2016) utilized NFI data as the primary source to generate information on carbon stocks (and thus emissions from forest change).

393 It is essential to understand not only carbon stocks in tropical forests through time but also the 394 response of tropical forest to climate change and develop policies accordingly. Information from 395 PSPs will allow us to determine whether Indonesian forests are sinks or sources of carbon and have 396 the potential to help us understand the factors driving carbon stock changes. To derive national 397 policies, information from PSPs needs to be combined with data on land use and land-use change, 398 which is accessible through remote sensing data or national inventories.

In addition, tropical forests are also key repositories of global biodiversity, genetic resources and important ecosystem services for local communities. Reducing biodiversity loss is a target of the United Nations Convention on Biological Diversity (Pereira et al. 2013) which is not only relevant from an aesthetic point of view, but can also threaten ecosystem functioning (Duffy 2009). Permanent sample plot data will foster a better understanding of the autecology, distribution and

- 404 rarity of tree species and they also have the potential to obtain measures of biodiversity of various
- 405 taxonomic groups at multiple scales and to link the abundances of each of these with one another.
- 406 All of the above are needed to enhance Indonesia's conservation planning efforts and manage forests
- 407 in a way that allows biodiversity to flourish in this exceptionally biodiverse country.
- 408
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- 412

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  Asia-Pacific Biodiversity Observation Network: Integrative Observations and Assessments. Springer, Tokyo, Japan.
  pp. 3-28

- 590 Table 1. Areas of forested land and sampled by permanent sample plots (PSPs) in primary
- 591 forest (excluding the National Forest Inventory) on major islands of Indonesia. Data on land
- 592 and forest area taken from Kementerian Lingkungan Hidup dan Kehutanan (2015b).

	Land area	Total forested	Primary forest area	Total PSP	PSP/forest
Island(s)	(10º ha)	<b>area</b> (10º ha)		<b>area</b> (ha)	area ratio**
			(10º ha)		
Java (+ Bali)	13.95	3.37	0.08	9.0	113.0
Sumatra	47.16	14.07	4.49	38.0	8.5
Kalimantan	52.96	27.58	9.80	82.1	8.4
Sulawesi	18.53	9.47	3.91	12.3	3.1
Nusa Tenggara*	6.76	2.84	0.68	2.5	3.7
Maluku	7.77	5.11	0.96	12.3	12.8
Papua	40.79	34.06	26.15	2.0	0.1
Total	187.92	96.50	46.07	158.1	3.4

593 \* Excluding Bali, which is included with Java due to their biogeographical affinity.

594 \*\* Area of permanent sampling plots (ha) per 10<sup>6</sup> ha of primary forest.

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- 598 Figure 1. (a) Locations of primary forest and (b) primary and disturbed permanent sampling
- 599 plots (PSPs) in Indonesia (excluding the National Forest Inventory).



Figure 2. (a) Plot areas, (b) total plot area under different lengths of monitoring and (c) total monitoring effort (i.e. sum of area multiplied by monitoring length for each plot) for permanent sample plots (PSPs) in primary forest (excluding the National Forest Inventory) on major islands of Indonesia. Note that plots only measured once are given a monitoring length of one year and also note the logarithmic scale for panel (c).



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