

Analyzing Transport Demand and Environmental Degradation: The Case of G-7 countries

Abstract

The debate for green development has been ongoing in the energy and environment literature—especially initiatives to mitigate climate change. On this note, we explore the effects of the air and railway transport demand, fossil-fuel energy consumption, demographic policies, economic growth, and alternative energy consumption on environmental degradation in Group of Seven (G7) economies. Using robust panel estimation techniques that account for cross-sectional dependency, empirical results affirm the presence of long-run relationships among variables. Besides, the results give credence to the environmental Kuznets Curve hypothesis (EKC) in G7 countries over the sampled period. We observe that demand for air transport, energy from fossil fuel sources, and economic development dampen environmental quality by 0.12%, 0.33%, and 46.54%, respectively. Interesting, renewable energy and rail transportation demand improves environmental quality. This outcome resonates with the need for alternative and clean energy production and consumption (Sustainable Development Goals (SDGs)-11 and 12) while improving the fight against climate change—especially the adoption of clean energy technologies in the air transport sector for sustainable growth.

JEL Codes: Q56; R4

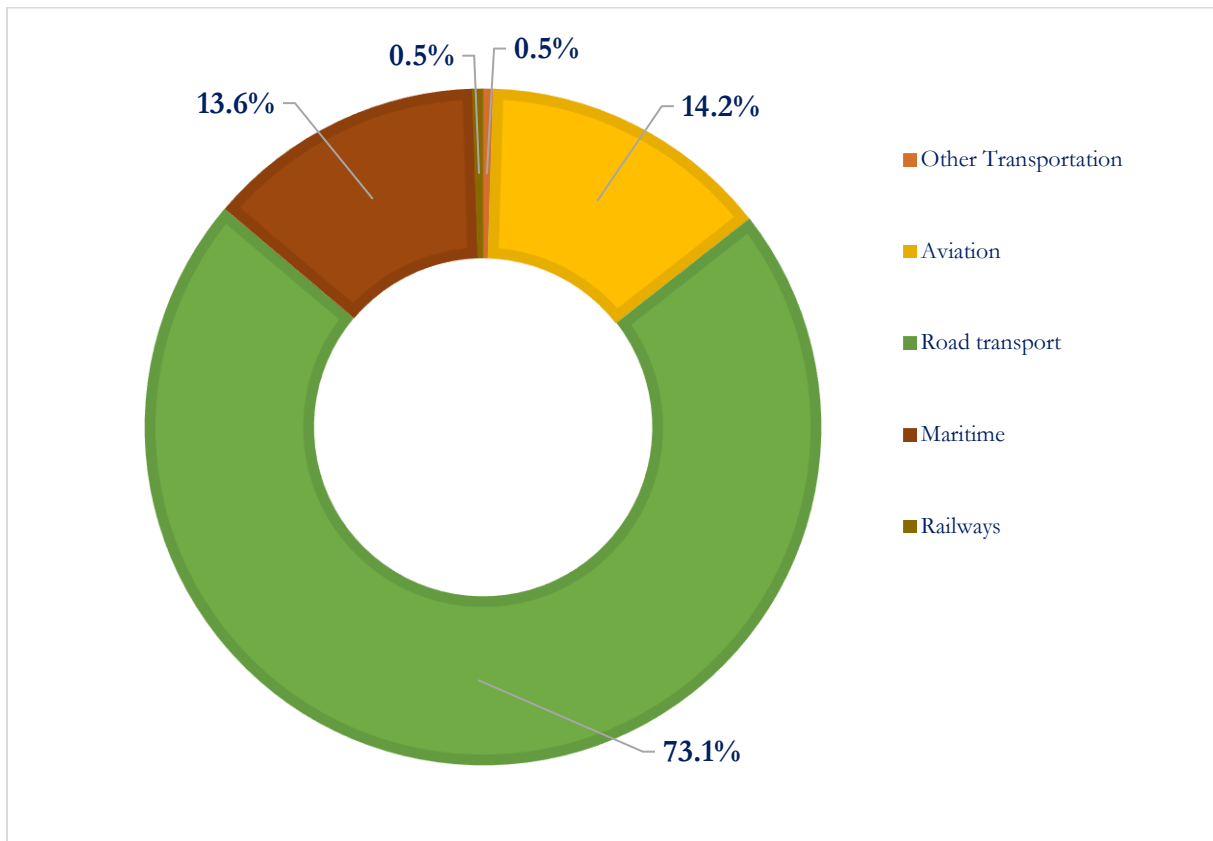
Keywords: Ecological footprint; heterogeneous effects; air transport; railway transport; EKC hypothesis; panel data modeling

24 **1. Introduction**

25 The last decade has witnessed an unprecedented increase in environmental research—which is
26 partly due to global concerns about environmental degradation and anthropogenic emissions. As a
27 result of the hazardous level that environmental degradation has reached today, many scholars—
28 typically in the environmental discipline—have performed different studies assessing various
29 factors including economic development (Aslan and Gozbasi 2016; Mikayilov et al. 2018; Adedoyin
30 et al. 2019), tourist activities (Rashid Khan et al. 2018), environmental Kuznets curve (EKC)
31 hypothesis (Esteve and Tamarit 2012; Aslan and Gozbasi 2016), renewable/fossil energy
32 consumption (Esso 2010; Esso and Keho 2016; Dogan and Ozturk 2017), which either upsurge or
33 decline emissions. These studies were conducted to highlight the policy direction on reducing
34 global anthropogenic emissions. According to the United Nations Global Climate Change Report
35 (2019), the current century is characterized by rising record levels of environmental degradation
36 and emission, thereby leading to toxic impacts on environmental quality. This means updating
37 policy directions requires consistent action, thus, assessing the linkage between travel demand,
38 energy consumption, and environmental degradation while introducing democracy and Kuznets
39 hypothesis, is a necessity for environmental policy development.

40 Land (i.e., road and railway), sea and air transport systems contribute significantly to the
41 global economy—nonetheless, amounts to the emanation of toxic substances such as carbon
42 dioxide emissions, methane, and others. According to the 2019 European Union for Climate
43 Action report, transportation is the major cause of pollution in European cities—accounting for
44 one-quarter of total greenhouse gas (GHG) emissions. As revealed in Figure 1, various forms of
45 transportation drive GHG emissions. Road transportation is the major contributor of GHG
46 emissions—accounting for 73.1%, followed by aviation or air transport—accounting for 14.2% of
47 total GHG emissions. Fossil fuel byproducts from aircraft business are riskier than other means of

48 transport activities—because it mutilates the arrangement of climate at a high range while
49 compounding GHG substance fixation at the surface climate. Thus, air transport increases the
50 toxicity of anthropogenic emissions, increasing environmental pollution. Besides, Chapman (2007)
51 posits that the 1997 Kyoto protocol emphasized the airline industry sector to decline global
52 anthropogenic emissions.



53
54

Figure 1. GHG emissions from the transportation sector.

55 Source: Authors' construction. Data Source: 2019 European Union Transport system.

56 Since the discharges from transportation present enormous danger to the environment and
57 sustained economic growth, the transport-emission nexus cannot be ignored—due to its role in
58 financial development. Consequently, there is a requirement for more firm procedures and
59 strategies on environmental policy to control the transport-emission nexus (Ouyang et al. 2019).
60 Hence, Grossman and Krueger (1995) re-explained the Kuznets Curve (KC) hypothesis proposed
61 by Simon Kuznets in 1995. According to this hypothesis, there exists an inverted U-shaped

62 association between economic growth per capita and income inequality. This implies economic
63 growth and income inequality simultaneously rise until a certain level—where the economy
64 continues to increase as income inequality begins to decline.

65 However, due to the continuous rise in global environmental issues, Grossman and
66 Krueger (1995) reinterpreted the curve as Environmental Kuznets Curve (EKC) hypothesis—
67 which explains the inverted U-shaped relationship between the environmental pressure and income
68 per capita growth. According to the EKC hypothesis, both income per capita and environment
69 degradation increase when income growth reaches a threshold—where environmental degradation
70 decreases thereafter. Stern (2004) argues that the factors of environmental emissions tend to
71 worsen due to advancement in economic development until a threshold of income level is achieved.
72 The EKC hypothesis is a useful tool to combat climatic problems caused by GHG emissions
73 (Sarkodie and Strezov 2019), thus, several empirical studies examine the validity of the hypothesis
74 (Esteve and Tamarit 2012; Aslan and Gozbasi 2016), and researchers employed several indicators
75 environmental degradation level such as sources of the energy, demographic indicators, technology,
76 source of economic activity, etc. (Sinha et al. 2019). However, less attention has been paid to
77 investigate the role of transport activities on environmental quality and a limited number of studies
78 investigated the role of transport activities on environmental degradation.

79 The primary research question of this paper is *“how do transportation activities interact with the*
80 *environment in G-7 countries?”*. Understanding the nature of the nexus between transportation choices
81 and environment can help to establish sustainable transport policies and to achieve environmental
82 and transportation aim of SDGs. To the best of our knowledge, this is the first paper attempts to
83 investigate which transport choice is more sustainable in G-7 countries. Therefore, the
84 investigation of the effect of transportation activities on the environment can provide deep insights
85 for decision-makers to establish efficient transportation policies that may help G-7 countries on
86 fulfilling their responsibilities arises from international environmental diplomacy. Moreover,
87 understanding the nature of the effect of transport choices may help policy-makers to affect

88 consumers behavior in long-run to guide their consumption patterns to more ecofriendly ones.
89 Second, we aim to contribute to improving existing literature by controlling omitted-variable bias.
90 Hence, we assess the effect of airline and railway transport demand alongside traditional variables
91 such as income, fossil fuels, and renewables on environmental degradation. Besides, we incorporate
92 democracy as the new factor to assess the role of institutional quality on long-term emission
93 reduction. The democratic system of governance may allow citizens to demand environmental
94 quality. Democratization or good institutional quality has social benefits/problems associated with
95 governments' systems of operation (Akalin and Erdogan 2021). Third, contrary to the adoption of
96 carbon emissions (CO₂) in existing literature, we use ecological footprint (EF)—more
97 comprehensive indicator that captures several dimensions of environmental accounting—to assess
98 long-term degradation effects. Ecological footprint captures the whole impact of human activities
99 on the environmental quality, it is required to consider multidimensional proxies of degradation
100 (Dasgupta et al. 2002; Kaika and Zervas 2013; Gill et al. 2018), so that genuine deterioration that
101 adversely affects the ecosystem and sustainable development would be comprehensively measured.
102 For this purpose, EF proposed by Wackernagel and Rees (1996) is taken as an alternative for CO₂
103 emissions in the context of the EKC hypothesis Fourth, we adopt novel estimation methods that
104 account for global common shocks, transboundary effects, and heterogeneous effects. Accounting
105 for such panel characteristics in the empirical model produces robust estimates useful for policy
106 formulation. Therefore, this paper combines three distinctive contributions to the literature on the
107 topic, sample and method.

108 The rest of the paper is organized as follows. Section 2 summarizes existing literature;
109 section 3 describes the used data and methodology and provides empirical results. Section 4
110 provides discussion of the empirical results. Finally, section 5 concludes the paper and gives the
111 policy recommendations.

112

113 2. Literature Review

114 **2.1 The air transport-induced-emissions nexus**

115 Burning a large percentage of oil is caused by the transport sector, one of the major
116 worldwide consumers of energy. This prompts the creation and discharge of carbon dioxide and
117 other ozone-depleting substances that may alter the climate, prompting environmental change.
118 Moreover, Rashid Khan et al. (2018) argued that the appropriate function of the air transport
119 system involves the utilization of a large amount of energy, thereby causing anthropogenic GHG
120 emissions that affect environmental quality. The energy burned from transportation can be linked
121 to travel demand among countries, thus, leading to tourism and tourist activities. So, reviewing the
122 empirical studies on tourism and emissions is tantamount to reviewing the transport-emission
123 nexus. In the lieu of this, several researchers have examined the transport-emission nexus while
124 suggesting possible ways of reducing emissions. For instance, Abdallah et al. (2013) revealed a
125 bidirectional relationship between transport and carbon emissions—implying CO₂ has a significant
126 impact on transport whereas transportation causes CO₂ emission. The overall effect of carbon
127 emissions on the environment affects air pollution while reducing economic growth, because of
128 spending required to reduce long-term pollution effects. Further research conducted by Gössling
129 et al. (2015) revealed increasing demand for air transport contributed to an upsurge in global climate
130 change from the airline sector. Thus, air travel contributes majorly to increasing ecological
131 degradation¹—due to the overdependence of the aviation sector on fossil fuels.

132

133 **2.2 Democracy, Institutional Quality, and Environmental Quality Nexus**

134 The relationship between democracy and environmental quality is a widely debatable issue
135 among scholars. Existing literature argues the role of democracy in reducing environmental
136 degradation, however, others assume democracy harms environmental quality. This section, thus,
137 presents the different theoretical and empirical arguments of the democracy-emission relationship.

¹ The toxic effects of fossil fuels have long-term environmental consequences

138 According to Payne (1995), the cause of environmental interest groups that leads to public
139 awareness and gather support for environmental enactment is encouraged by the opportunity for
140 individuals to exercise political rights and freedom of information. This works through public
141 opinions on environmental issues, thus, information on ecological issues streams more openly, and
142 political rights are more preferably ensured in a vote-based system over totalitarianism.
143 Additionally, Magnani (2000) proposes that property rights, democracies, and regard for basic
144 liberties can create synergies that lead to expanded levels and adequacy of environmental strategy.

145 Ecological gatherings are frequently more fruitful at educating individuals and coordinating
146 them to follow-up on environmental issues in a democracy than in a despotism (Schultz and
147 Crockett 1990). Another argument is that institutional quality may likely promote environmental
148 quality in a democratic system of governance than an autocratic system. This is because
149 democracies are bound to follow environmental arrangements since they regard the standard of
150 law, thus, raises ecological quality (Weiss and Jacobsen 1999). According to Berge (1994), monetary
151 opportunity is regarded by democracies, thus, market economies promote ecological quality.
152 Gleditsch and Sverdrup (2003) propose that as vote-based systems regard human existence more
153 than despotism, they are more receptive to perilous ecological debasement. Likewise, vote-based
154 systems participate in fewer wars, hence, have a more significant level of ecological quality since
155 wars obliterate the climate.

156 Several components rely on the view that environmental degradation may or may not lessen
157 by democracies. Hardin (1968) cautions about the approaching danger of unchecked and ecological
158 mismanagement without support from concerned policymakers. At the point when private
159 property privileges of regular assets are not well-defined, people or environmental groups tend to
160 overuse such resources and overlook the harm that monetary activities incur on the climate.
161 Additionally, Paehlke (1996: 28) argues that the global economy and climate pose extraordinary
162 peril for democracy and the environment. This is because the functionality of democracy is based
163 on national levels, thus, global policy directions on environmental issues may not be topical for

164 discussion in time. Heilbrunner (1974) contends that the increase in global population growth
165 contributes to the menace from environmental degradation. However, since democracies respect
166 citizens' right to live and reproduce, imposing action to curtail human multiplication will be
167 difficult, unlike autocracies whose leaders are autonomous decision-makers.

168 Where ecological corruption is concerned, democracies regularly experience public strategy
169 inaction. Democratic leaders tend to please the contending interest parties in the public to win
170 votes (Midlarsky 1998). Also, there can be a long-time argument between corporates and
171 environmental groups, thereby ignoring a democratic decision on the action that will advance
172 environmental quality. Spending limitation may hinder the responsiveness of government to
173 ecological imperatives but to additional major problems of financial means of significant parts of
174 the democratic public (Midlarsky 1998: 159). Moreover, democracies are hesitant to ease natural
175 corruption since certain gatherings rely on profit (or loss) from ecological strategies more than
176 others (Midlarsky 1998: 159).

177 The environmental degradation and democracy nexus has long-term being the inconclusive
178 subject of controversies among economists, political theorists, natural resource management
179 studies, and social geography. Some scholars refuse to admit that environmental disruption is not
180 explained by the importance of democracies. Roberts and Parks (2007) modeled democracy as a
181 determinant of environmental pollution, proxied by CO₂ emission, and found democracy has no
182 significant effects on CO₂ emission. Scruggs (1998) used three different environmental proxies
183 while controlling for economic inequalities, and found no significant relation with democracy.
184 Among scholars who reveal significant associations between democratic levels and environmental
185 quality are Payne (1995), Barrett and Graddy (2000), Farzin and Bond (2006), and Torras and Boyce
186 (1998). These scholars admit that people are more resilient in seeking environmental quality in a
187 democratic system because of early warnings, and environmental awareness, hence, become more
188 receptive to requests for environmental protection.

189

190 **2.3 Other determinants of Environmental Quality**

191 The main policy objectives of global communities are to control the toxic impact of
192 environmental change and lessen global GHG emissions. This responsibility relies more on major
193 polluters of CO₂ to meet specific emission targets outlined in both the Kyoto Protocol and Paris
194 agreement. Thus, understanding the fundamental factors of emissions is useful in developing
195 policy-oriented towards climate change reduction. Because of the obvious effects of environmental
196 change on economic development and quality of life, several studies for decades have investigated
197 the connection between ecological degradation, economic factors (such as GDP, EKC hypothesis,
198 FDI, and travel demand), political factors, institutional factors, and energy consumption. For
199 instance, Yoo (2006) found a two-way causal link between the consumption of coal rents and
200 economic growth. This validates a mutual effect between economic growth and coal consumption,
201 thus, as economic growth booms, more coal resources are exploited and consumed, hence,
202 disrupting air quality. Also, Li and Li (2011) found a one-directional causal link from GDP per
203 capita to coal consumption in selected countries whereas the opposite effect was confirmed in
204 other countries. In both scenarios, the GDP and coal consumption relationship degrade
205 environmental quality, confirming a similar study by Apergis and Payne (2010).

206 Interest in improving economic growth and social investment has soared over decades, but
207 the increment without sustainability would require burning more fossil fuels, thereby releasing CO₂,
208 a major cause of GHG emissions that leads to climate change. This mystery brought the EKC
209 theory that examines the relationship between economic growth (GDP per capita) and GHG
210 emissions. For instance, Bagliani et al. (2008) revealed that the increase in economic growth does
211 not affect environmental degradation. The same result is confirmed in Caviglia-Harris et al. (2009)
212 and Wang et al. (2013). In another case study, Odhiambo (2012) averred the existence of a one-
213 way causal relationship from GDP per capita to carbon emission whereas energy consumption
214 predicts emissions and economic development. This statement means economic productivity spur
215 emissions, but a feedback effect is observed between energy consumption and emissions. Similarly,

216 energy is needed for economic development, hence, energy demand increases as economic
 217 activities grow rapidly. In contrast, a study with a panel of Organisation for Economic Co-
 218 operation and Development (OECD) and non-OECD countries revealed that while emissions
 219 underpin economic productivity in OECD countries, emissions do not trigger economic growth
 220 in non-OECD countries (Dinda 2009). Several studies in Table 1 have assessed the validity of the
 221 EKC hypothesis—which explains an inverse U-shaped relationship between economic growth and
 222 emissions. The empirical review in Table 1 is an investigation of different proxies of environmental
 223 degradation, and income. The last column of Table 1 shows the inconclusive validity of the EKC
 224 framework across several studies. It is observed that the same countries and regions show
 225 disparities in empirical results, showing the complexity of the EKC hypothesis. For instance, Ajmi
 226 et al. (2015) and Ari and Senturk (2020) find no evidence of the EKC hypothesis for G-7 countries
 227 whereas Nabaee et al. (2015), Raza and Shah (2018), Answer et al. (2020), and Wang et al. (2020)
 228 support the validity of the EKC hypothesis. Contrary, Shahbaz et al. (2017) find mixed results of
 229 the EKC hypothesis across selected countries.

230

231

Table 1. Validity of the EKC hypothesis

S/N	Authors	Country	Methods	Validity of EKC
1.	Halicioglu (2009)	Turkey	ADF, ARDL, VECM causality	No EKC
2.	Nasir and Rehman (2011)	Pakistan	ADF, PP, Johansen cointegration	EKC
3.	Jalil and Faridun (2011)	BRIC	LLC, IPS, ADF, PP, OLS, VECM	EKC
4.	Ajmi et al. (2015)	G7	time-varying Granger causality analysis	No EKC
5.	Nabaee et al. (2015)	G7	panel cointegration tests	EKC
6.	Shahbaz et al. (2017)	G7	nonparametric time series	Mixed
7.	Raza and Shah (2018)	G7	FMOLS, DOLS, FE-OLS	EKC

S/N	Authors	Country	Methods	Validity of EKC
8.	Anser et al. (2020)	G7	panel random effects,	EKC
9.	Ari and Senturk (2020)	G7	panel cointegration tests and long-term estimators	No EKC
10.	Wang et al. (2020)	G7	dynamic seemingly unrelated regression	EKC
11	Ahmad et al. (2021)	G7	CS-ARDL	EKC
12	Pata and Caglar (2021)	China	Augmented ARDL	No EKC
13	Qin et al. (2021)	G7	CS- ARDL	EKC
14	Balsalobre-Lorente et al. (2021)	EU-5	panel cointegration tests, FMOLS	EKC
15	Frodyma et al. (2022)	EU	ARDL-bounds testing	No EKC
16	Çakar et al. (2022)	G7	Panel Threshold Regression	No EKC
17	Doğan et al. (2022)	G7	Panel cointegration tests, FMOLS, DOLS	EKC
18	Zhao et al. (2022)	G7	CCEMG, AMG	EKC
19	Miao et al. (2022)	NICs	Method of Moments Quantile Regression (MMQR)	EKC
20	Zhao et al. (2022)	G7	CS-ARDL	EKC

232

233 3. Data, Methodology, and Empirical Results

234 3.1. Data and Methodology

235 The economic development-environmental quality nexus has long been a significant discussion for
236 policymakers and researchers since the original study of Grossman and Krueger (1991). The model
237 specification became one of the significant pathways in subsequent empirical studies. On one hand,

238 researchers have focused on the issue of the choice of proxy variables for environmental
239 degradation. Grossman and Krueger (1991) employed carbon emissions per capita as an indicator
240 of environmental pollution, hence, many studies adopted a similar modeling strategy. However,
241 the adoption of carbon emissions per capita has been criticized because of its one-dimensional
242 environmental degradation tendency (Destek et al. 2018; Erdogan and Okumus 2020). Contrary,
243 some studies have since utilized ecological footprint per capita proposed by Wackernagel and Rees
244 (1998) because of its embracive proxy of environmental pollution (Solarin 2019). Ulucak and Lin
245 (2017) asserted that ecological footprint considers anthropogenic pressure on the ecological system
246 on six sub-component including footprints of carbon, forest, grazing land, cropland, fishing
247 ground, and built-up land, therefore, it could be regarded as a multidimensional indicator for
248 environmental pollution compared to CO₂ emissions. On the other hand, Grossman and Krueger
249 (1991) employed international trade to increase the robustness of the EKC model, and international
250 trade has become one of the significant parts of the empirical models. Afterward, sub-components
251 of energy consumption have been integrated into empirical models. Energy-related regressors are
252 essential parts of modeling environmental pollution because energy is considered the main driver
253 of environmental deterioration (Sinha et al. 2019).

254 Furthermore, Hillman and Ursprung (1992) emphasized that the political system influences the
255 environmental performance of countries. Median voters directly express their policy choice in
256 direct democracies, which hinders policy-makers to adopt discretion-based environmental policies.
257 Contrary, policymakers can adopt environmental policies against the majority of voters in
258 representative democracies, and monitoring discretion-based environmental policies could be
259 difficult for voters because of the principal-agent problem. In such a case, interest groups could
260 use the lack of effective monitoring, hence, governments may take environmental to please
261 lobbying groups instead of the best interests of voters (Akalin and Erdogan 2020; Dryzek 1987;
262 Olson 1982). Besides, autocratic governments may object to enfranchising the demands of
263 lobbying groups on ecological issues and prioritize public interest over lobbying groups. Thus,

264 autocratic regimes can exhibit better performance in the implementation of international-based
265 environmental commitments because of long administration terms (Bernauer and Koubi 2009).
266 Moreover, individuals have a greater opportunity to have information on environmental policies
267 and freely express their policy choices in democracies than in autocracies. In this manner,
268 individuals can put political pressure on politicians by using democratic institutions to have a
269 cleaner and sustainable environment (Acemoglu and Robinson 2006).

270 A significant segment of the former studies used economic activity-based indicators including
271 industrial and agricultural production, energy production, transportation activities, and forestry.
272 Transportation activities ease the mobility of labor, goods, and services—strengthen competition
273 and market mechanism—reduces distribution cost, and limit monopolistic structures. Individuals
274 can easily access health services and attend educational activities through transportation activities
275 (Erdogan 2020). However, access to transportation is not without cost. The Intergovernmental
276 Panel on Climate Change (2014) (IPCC) reported that transportation and its sub-sectors accounted
277 for approximately 14% of total anthropogenic emissions. Muñoz-Villamizar et al. (2020) reported
278 that an increase in urban population would inevitably lead to an increase in transportation demand.
279 The International Energy Agency (2020) reported that environmental pollution may tend to
280 increase due to increasing transportation demands. In this regard, reducing transport-led
281 environmental pollution is one of the significant targets of international treaties. For instance, the
282 Kyoto Treaty highlighted transport-originated pollution as one of the significant environmental
283 burdens for countries (Chapman 2007). Besides, great significance has been attributed to meet the
284 demand for accessible, safe, affordable, and sustainable transportation in the Sustainable
285 Development Goals (SDGs). It is emphasized that transportation activities would be expected to
286 be the major driving force behind a growing energy demand particularly in developed countries—
287 which in turn spur environmental degradation (Erdogan 2020; Georgatzi et al. 2020; United
288 Nations 2020). Thus, efficient transport demand management will help to achieve both climate
289 action (SDG-13) and sustainable transport (SDG-11.2).

290 Based on this theoretical framework, we employed the following model to investigate the nexus
291 between transport demand and environment nexus in G-7 countries for the period spanning 1995
292 to 2015:

$$293 \ln EF_{it} = \beta_{0i} + \beta_1 \ln Y_{it} + \beta_2 \ln Y_{it}^2 + \beta_3 \ln AR_{it} + \beta_4 \ln RL_{it} + \beta_5 \ln D_{it} + \beta_6 \ln F_{it} + \beta_7 \ln R_{it} + \varepsilon_{it} \quad (1)$$

294 Where EF represents ecological footprint per capita, Y is the gross domestic product (GDP) per
295 capita (constant, 2010 US\$), while Y_{it}^2 is GDP per capita square. The AR is airline transport
296 demand (number of passengers (carried)), RL is the railway transport demand (number of
297 passengers (carried)). AR and RL were employed as a proxy of air and railway transport demand.
298 D is the democracy index, F is the fossil fuel energy consumption (% of the total), and R is
299 renewable energy consumption (% of total final energy consumption). Data for ecological footprint
300 were obtained from the Global Footprint Network (2019) whereas data for GDP per capita (Y),
301 fossil fuel energy consumption (F), and renewable energy consumption (R) were retrieved from
302 World Development Indicators published by the World Bank (2020). We formed the democracy
303 employing two different sub-components and it is a composite index of law-order and democratic
304 accountability estimated by taking the simple average of two indicators as an indicator of
305 democracy. The law-order and democratic accountability indexes vary between 0 and 6, where 0
306 shows poor performance, while 6 shows high performance on democracy. The law-order and
307 democratic accountability indexes were retrieved from the International Country Risk Guide
308 (ICRG) (2020). All variables were converted to logarithmic values by taking natural logarithm, and
309 empirical estimations were conducted by using EViews 10, Stata 15 and Gauss 17.

310 Contrary to the static model presented in equation (1), we re-estimated the proposed model using
311 a dynamic panel technique that controls for heterogeneous effects, omitted-variable bias, hence,
312 produces consistent and robust parameters. Following the empirical procedure presented in

313 Sarkodie and Owusu (2020); Owusu and Sarkodie (2020), we adopt a dynamic panel with
 314 bootstrap-corrected fixed-effects expressed as:

$$\begin{aligned}
 315 \quad \ln EF_{i,t} &= \lambda \cdot \ln EF_{i,t-1} + \beta_1 \cdot \ln Y_{i,t} + \beta_2 \cdot \ln Y_{i,t}^2 + \beta_3 \cdot \ln AR_{i,t} + \beta_4 \cdot \ln RL_{i,t} + \beta_5 \cdot \\
 316 \quad \ln D_{i,t} &+ \beta_6 \cdot \ln F_{i,t} + \beta_7 \cdot \ln R_{i,t} + \delta_{i,t} + \varepsilon_{i,t} \qquad (2)
 \end{aligned}$$

317 Where, $\ln EF_{i,t}$ is the target variable, $\ln EF_{i,t-1}$ is the lagged-dependent variable with autoregressive
 318 coefficient $\lambda < 1$ —signifying a dynamically stable relationship between ecological footprint and
 319 regressors [$\ln Y_{i,t}$, $\ln Y_{i,t}^2$, $\ln AR_{i,t}$, $\ln RL_{i,t}$, $\ln D_{i,t}$, $\ln F_{i,t}$, and $\ln R_{i,t}$], β_1, \dots, β_7 denote the estimated
 320 parameters performed with parametric bootstrap technique, $\delta_{i,t}$ denotes unobserved
 321 heterogeneous effects across countries while accounting for country-specific fixed-effects, and $\varepsilon_{i,t}$
 322 is the error term across cross-sections $i=1, \dots, N$ and period $t=2, \dots, T$. The model specification
 323 in equation (2) is estimated using both cross-sectional heteroskedasticity and contemporaneous
 324 cross-sectional dependence for resampling of the error term—with either burn-in or analytical
 325 heterogeneous initializations. Both standard errors and confidence intervals are estimated using the
 326 fixed-effects bootstrap distribution estimator (De Vos et al. 2015).

327 To investigate cross-section dependence among variables and model, we implemented a cross-
 328 section dependence (CD) test developed by Pesaran (2004). We unearthed the stationarity
 329 properties of sampled variables—which is cross-sectionally independent—by using Im-Pesaran-
 330 Shin (IPS) unit root test proposed by Im et al. (2003) to investigate the unit root properties. The
 331 IPS estimation can be carried out by using Eq. (3):

$$332 \quad \Delta y_{it} = \alpha_i + \beta_i y_{i,t} + \varepsilon_{it} \qquad (3)$$

333 Where α_i and β_i extended as $(1-\phi_i)\mu_i$ and $-1(1-\phi_i)$, respectively. $\Delta y_{it} = y_{it} - y_{i,t-1}$. The
 334 IPS test adopts “non-stationarity” in the null against “stationarity” in the alternative. Moreover, we
 335 utilized the cross-sectionally augmented IPS (CIPS) test proposed by Pesaran (2007) to investigate

336 the stationarity of cross-sectionally dependent variables. The CIPS test can be implemented by
 337 taking simple averages of country-specific estimations: $CIPS = \sum_{i=1}^N CADF_i / N$. The CIPS method
 338 adopts “non-stationarity” in the null against “stationarity” in the alternative. Due to non-normal
 339 distributions, critical values of the CIPS test for different sample sizes are estimated by Monte
 340 Carlo methods.

341 After determining the integration level of the variables, we conducted Pedroni (1999) method to
 342 examine the existence of possible cointegration in the model. The Pedroni cointegration form of
 343 Eq. (1) can be estimated by using the following specification:

$$344 \quad \varepsilon_{it} = \psi_i \varepsilon_{it-1} + \sum_{k=1}^{K_i} \psi_{ik} \varepsilon_{it-k} + v_{it} \quad (4)$$

345 Pedroni's (1999) test adopts “no-cointegration” in the null against “cointegration” in the
 346 alternative. To estimate long-run coefficients, we used the Fully Modified Ordinary Least Squares
 347 (FMOLS) approach developed by Pedroni (2000). The panel specification of the FMOLS method
 348 can be applied as (Erdogan 2020):

$$349 \quad \beta_{GMOLS} = N^{-1} \sum_{i=1}^N \beta_{FMOLSi} \quad (5)$$

350 Where β_{FMOLSi} is computed by employing country specific FMOLS computation of Eq. (1). The
 351 t-statistics for the coefficients can be computed as: $t_{\beta_{GMOLS}} = N^{-1/2} \sum_{i=1}^N t_{\beta_{FMOLSi}}$.

352

353 **3.2. Empirical Results**

354 We began our analysis by implementing a CD test to examine the effect of global common shocks—
 355—in this way, the potentiality of biased estimation is hindered, producing robust parameters for
 356 policy proposals. The CD test results (Table 2) show the null hypothesis of “no cross-section
 357 dependence” is accepted in railway transport demand, whereas the null hypothesis of cross-section

358 dependence is accepted for other variables. Hence, traditional panel unit root techniques can be
 359 implemented for railway transport demand, whereas panel unit root tests that perform well under
 360 cross-section dependence are feasible for the remaining variables. Moreover, the null hypothesis
 361 of “no cross-section dependence” is accepted for the cointegration model, thus, first-generation
 362 panel cointegration tests and parameter estimation methods can be implemented (Erdogan et al.
 363 2020b).

364 **Table 2.** CD Test Results

Test	MODEL	EF	Y	Y ²	AR	RL	D	F	R
CD -Stat.	1.350	13.610	16.200	16.190	12.370	1.330	4.54	5.380	15.960
	(0.176)	(0.000)	(0.000)	(0.000)	(0.000)	(0.180)	(0.000)	(0.000)	(0.000)

365 **Legend:** *EF* represents ecological footprint per capita, *Y* is the gross domestic product (GDP) per capita (constant,
 366 2010 US\$), *Y*² is GDP per capita square, *AR* is The *AR* is airline transport demand (number of passengers (carried)),
 367 *RL* is the railway transport demand (number of passengers (carried), *D* is the democracy index, *F* is the fossil fuel
 368 energy consumption (% of the total), and *R* is renewable energy consumption (% of total final energy consumption)).

369
 370 To determine stationarity levels of railway transport demand, we used the IPS test as a first-
 371 generation unit root method. The findings (Table 3) show that the null hypothesis of “unit root”
 372 is accepted at a 5% significance level for railway transport demand in level, whereas it turns
 373 stationary at a 5% significance level in the first difference.

374

375 **Table 3.** IPS Results

Indicator	Model 1	Model 2
	c+t	c+t
	\bar{t}	\bar{t}
<i>RL</i>	-1.326	-3.722

(0.923)

(0.000)

376 **Note:** Maximum lag length was used as $k=2$. Probability values of the IPS test were reported in parenthesis. Model
377 1: Level, Model 2: 1st Difference, c: constant, t: trend.

378 After determining the integration levels of the cross-sectionally independent variable, we
379 implemented the CIPS test to investigate the integration level of the cross-sectionally dependent
380 variables. The CIPS results in Table 4 reveal ecological footprint, income, income square, airway
381 transport demand, democracy, fossil energy, and renewable energy consumption follow the unit
382 root process at 5% significance levels in level. However, the first difference of such variables
383 follows the stationary process at a 5% significance level.

384 **Table 4.** CIPS Results

Indicator	Model 1	Model 2
	c+t	c+t
	\bar{t}	\bar{t}
EF	-2.883	-4.256
Y	-1.303	-3.157
Y ²	-1.906	-3.153
AR	-2.332	-4.896
D	-2.617	-4.096
F	-1.477	-5.043
R	-2.344	-5.327

385 **Note:** Maximum lag length was used as $k=2$. Model 1: Level, Model 2: 1st Difference, c: constant, t: trend. Critical
386 values for CIPS method are -2.76 (10%), -2.94 (5%), and -3.30 (1%), respectively. **Legend:** EF represents ecological
387 footprint per capita, Y is the gross domestic product (GDP) per capita (constant, 2010 US\$), Y² is GDP per capita
388 square, AR is The AR is airline transport demand (number of passengers (carried)), RL is the railway transport
389 demand (number of passengers (carried)), D is the democracy index, F is the fossil fuel energy consumption (% of the
390 total), and R is renewable energy consumption (% of total final energy consumption)).

391 The possibility of the emergence of cointegration nexus among variables occurs in the case of
392 integrational levels at $I(1)$. Therefore, we implemented the Pedroni approach to investigate whether

393 the variables are cointegrated in the long run. The Pedroni test findings (Table 5) confirm the
 394 rejection of the null hypothesis of “no cointegration” at a 5% significance level in three different
 395 statistics—therefore, income, the second-degree polynomial of income, air and railway transport
 396 demand, democracy, fossil, and renewable energy have a long-run effect on ecological footprint.
 397 Next, we assessed heterogeneous effects across countries using the novel panel kernel-based
 398 heterogeneous test with output plots presented in Figures 2 and 3. Evidence from Figures 2 and 3
 399 reveals the mean point estimates of data series are within the 95% confidence interval, thus,
 400 confirming heterogeneous effects across countries.

401

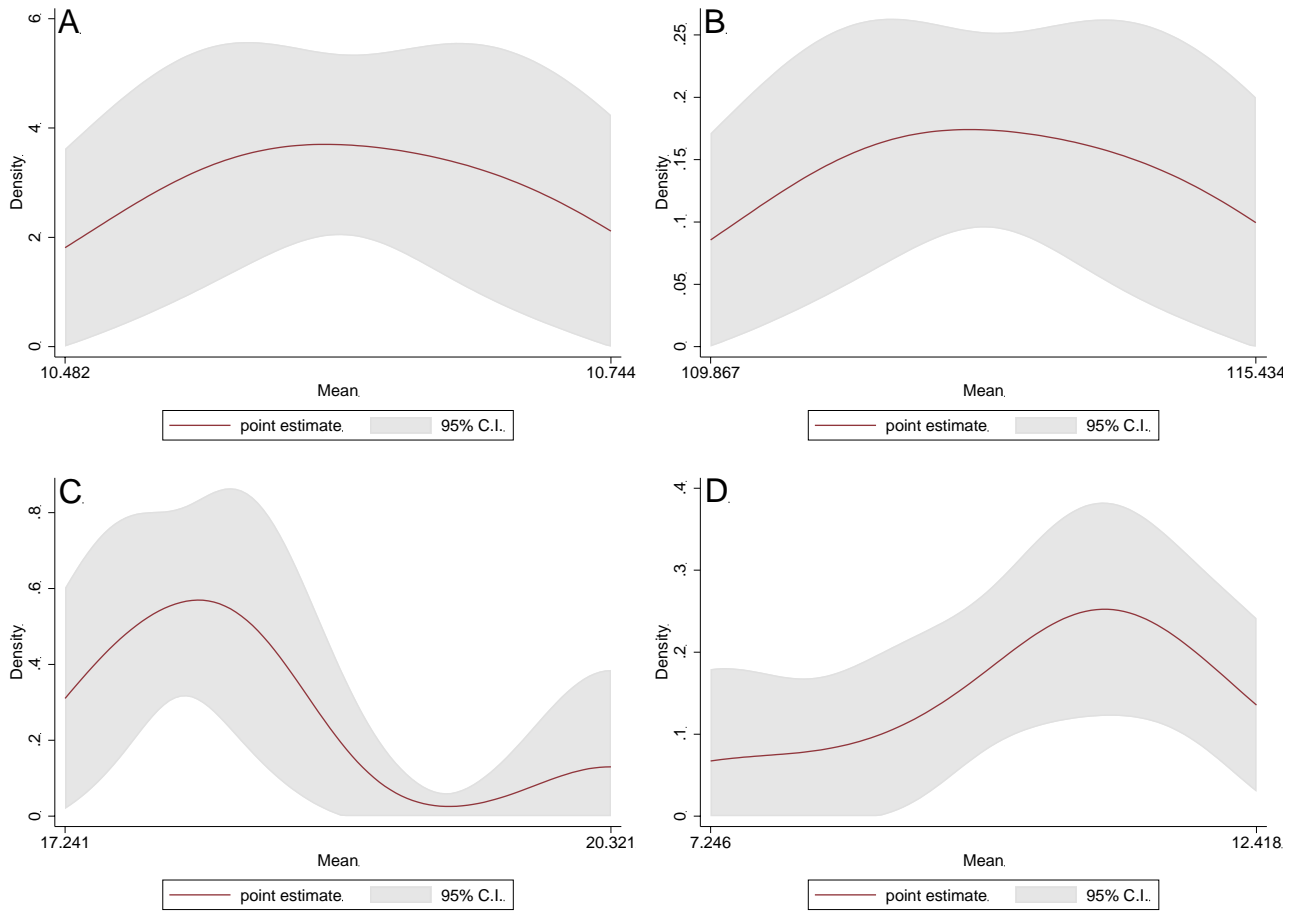
402 **Table 5.** Pedroni (1999) Test Results

	Stat.
Modified PP	2.830 (0.002)
PP	-5.364 (0.000)
ADF	-5.024 (0.000)

403 **Note:** Probability values for the Pedroni cointegration test are provided in the parenthesis. PP: Phillips-Perron,

404 ADF: Augmented Dickey-Fuller.

405



406

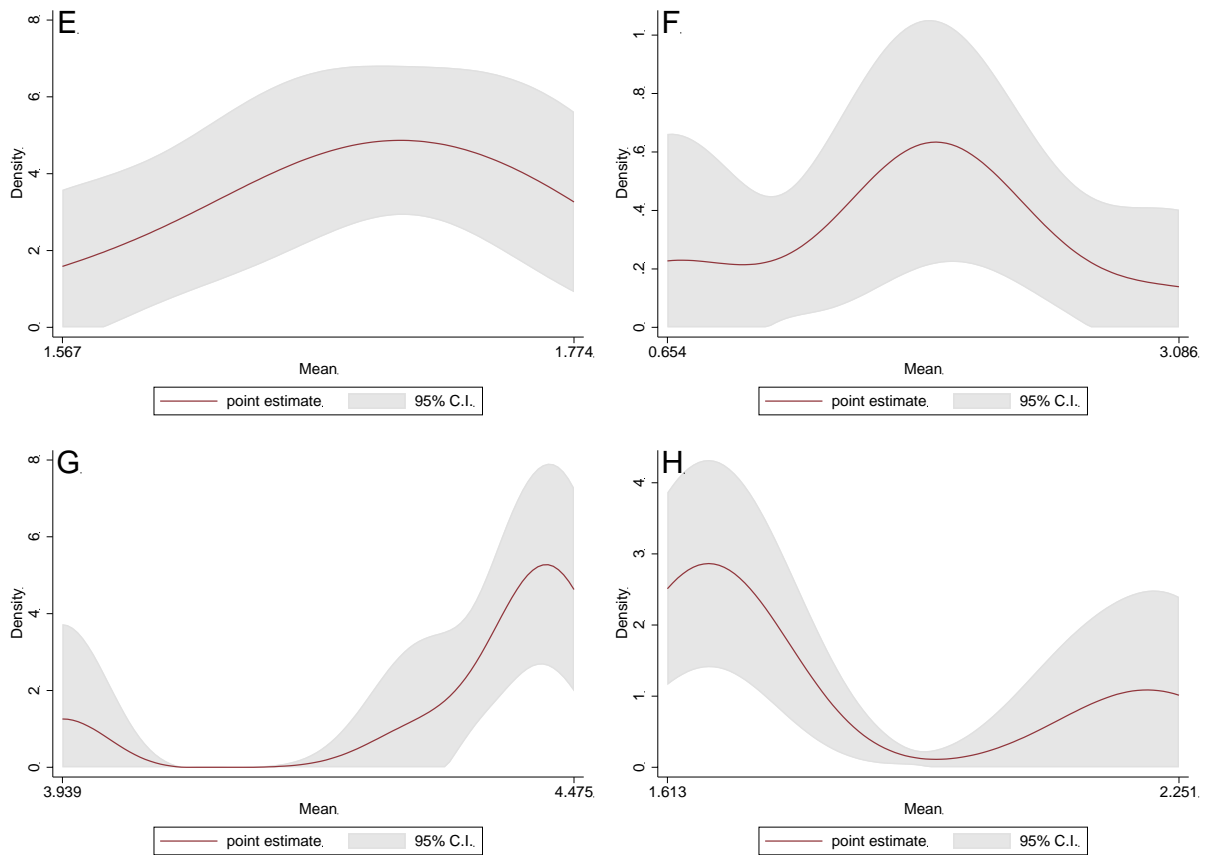
407 Figure 2. Panel kernel-based heterogeneous effects (A) GDP (B) GDP^2 (C) Airline transport

408

demand (D) Railway transport demand.

409

410



411

412 Figure 3. Panel kernel-based heterogeneous effects (E) Democracy index (F) Renewable energy
 413 consumption (G) Fossil fuel energy consumption (H) Ecological footprint per capita.

414

415 The long-run coefficient estimation results using FMOLS specification are shown in Table 6.
 416 Accordingly, income positively affects ecological footprint (i.e., at $p\text{-value} < 0.05$) whereas income
 417 square negatively affects ecological footprint (i.e., at $p\text{-value} < 0.05$). These results validate the
 418 existence of the EKC hypothesis. The empirical results infer poor environmental quality at the
 419 initial economic pathway in G-7 countries, however, environmental quality improved after passing
 420 a certain level of income, viz. the turning point of inverted U-shaped curve. In other words,
 421 environmental degradation tends to increase in the first phase of the EKC while degradation
 422 declines at the second stage of the EKC. These findings confirm one part of existing literature (see
 423 Erdogan et al. (2020a); Acaravci and Akalin (2017); Sarkodie and Adams (2018); while in
 424 contradiction with another set of existing literature (see Aslan et al. (2018); Ozcan et al. (2018)

425 Sarkodie and Strezov (2019)). Moreover, we computed a turning point of income as US\$ 45,738
 426 per capita. In this regard, Canada, Japan, and the US have already passed the threshold level of
 427 income, whereas France, Germany, Italy, and United Kingdom are yet to achieve such a target (i.e.,
 428 level of income in 2015).

429 **Table 6.** FMOLS Estimation Results

Indicators	Coefficients	t-stat.	Prob.
Y	46.535	1250.074	0.000
Y^2	-2.168	-39.631	0.000
AR	0.121	2.054	0.042
RL	-0.003	-0.061	0.950
D	0.182	5.095	0.000
F	0.332	13.301	0.000
R	-0.178	-4.756	0.000

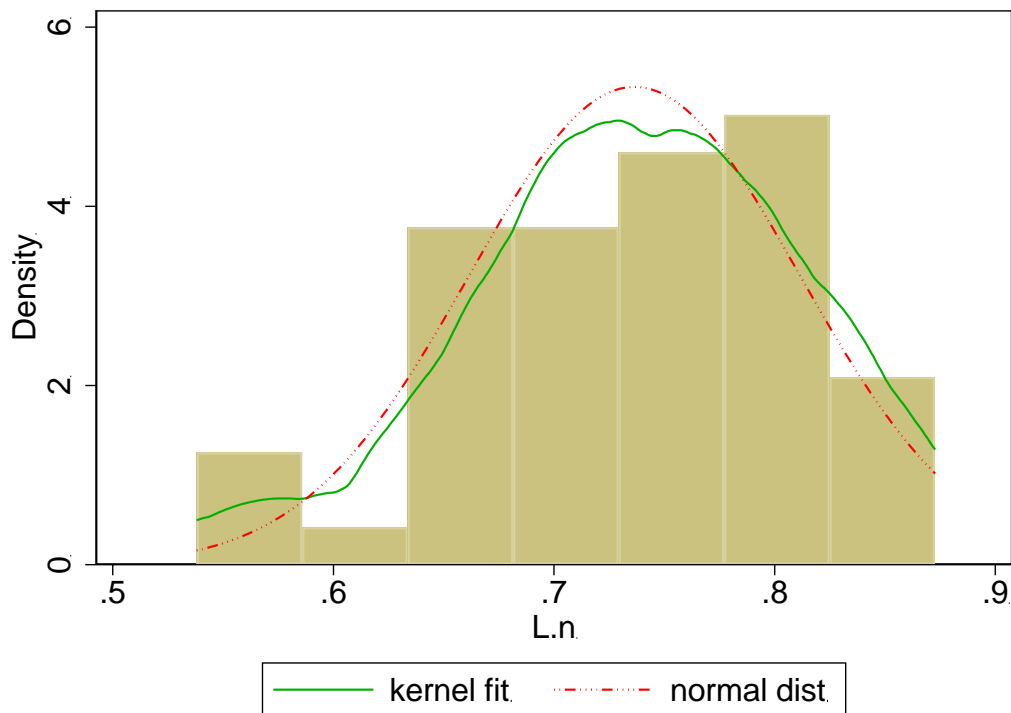
430 **Note:** The covariance for the long-run was computed by using the Newey-West fixed bandwidth and Bartlett kernel
 431 based on the heterogeneity of long-run variance. **Legend:** EF represents ecological footprint per capita, Y is the
 432 gross domestic product (GDP) per capita (constant, 2010 US\$), Y^2 is GDP per capita square, AR is The AR is airline
 433 transport demand (number of passengers (carried)), RL is the railway transport demand (number of passengers
 434 (carried), D is the democracy index, F is the fossil fuel energy consumption (% of the total), and R is renewable
 435 energy consumption (% of total final energy consumption)).

436
 437 Moreover, a 1% increase in air transport demand increases ecological footprint by 0.12% (i.e., at p -
 438 $value < 0.05$), hence, air transport demand positively affects environmental degradation. This implies
 439 a positive shock in travel demands by air transportation may lead to a rise in environmental
 440 pollution levels. These findings confirm the results of Erdogan (2020); Erdogan et al. (2020a), while
 441 contradicting Pereira and Pereira (2017). The railway transport demand has a negative but
 442 statistically insignificant effect (i.e., at p - $value > 0.05$) on ecological footprint. Thus, we confirm no
 443 statistically significant nexus between railway transport demand and ecological footprint. This is

444 consistent with Georgatzi et al. (2020); Neves et al. (2017), while in contrast with Erdogan (2020);
445 Erdogan et al. (2020a); Pereira and Pereira (2017). Besides, 1% increase in democracy escalates
446 ecological footprint by 0.18% (i.e., at $p\text{-value}<0.05$). Democracy has a positive and statistically
447 significant effect on ecological footprint—implying democratization exacerbates the
448 environmental condition in G-7 countries. This supports the findings of Akalin and Erdogan
449 (2020); Mak Arvin and Lew (2011); Charfeddine and Mrabet (2017); while in contrast with Bernauer
450 and Koubi (2009); Pellegrini and Gerlagh (2006); Adams and Klobodu (2017). Fossil energy use
451 has a positive and statistically significant effect on ecological footprint—consequently, a 1%
452 increase in fossil fuel consumption increases ecological footprint by 0.33% (i.e., at $p\text{-value}<0.05$). In
453 contrast, a 1% increase in renewable energy utilization decreases ecological footprint by 0.17% (i.e.,
454 at $p\text{-value}<0.05$), inferring the adoption of renewables safeguard ecological deterioration. The
455 findings related to both fossil fuels and renewables are consistent with several existing literature.
456 The larger coefficient of fossil fuels shows the mitigating effect of renewables is yet to offset the
457 deteriorating effect of fossil fuel energy consumption. This segment of empirical results is parallel
458 with theoretical expectations and former literature (see Lv (2017); Farzanegan and Markwardt
459 (2018); Kim et al. (2019)).

460 We further validate the FMOLS estimator using a panel dynamic bootstrapped-corrected estimator
461 with country-specific fixed effects. To further verify the estimated dynamic model, we followed
462 the testing procedure expounded in Owusu and Sarkodie (2020) to examine the distributional
463 structure of the estimated residuals (Figure 4). We employed the bootstrap distribution for
464 diagnostic analysis—displayed in histogram alongside kernel fit and normal distribution plots
465 depicted in Figure 4. The bootstrap distribution dynamics of the estimated residuals—designated
466 by histogram plot presented in Figure 4—shows a kernel fit that follows the normal distribution.
467 The bootstrap distribution of the sum of autoregressive parameters reveals a seemingly perfect
468 prediction of the estimated model that validates the robustness of the model specification and

469 statistical inferences. Thus, confirming the residual independence of the estimated dynamic model
 470 achieving convergence and stability over time and across sampled countries.



471
 472 Figure 4. Bootstrap distribution dynamics of GDP, GDP², airline transport demand, railway
 473 transport demand, democracy index, renewable energy consumption, and fossil fuel energy
 474 consumption in ecological footprint function.

475
 476 The subsequent results of the estimated model are presented in Table 7. Similar signs of FMOLS
 477 coefficients are observed in the heterogeneous panel estimation results, however, the level of
 478 statistical significance differs because of the lagged-dependent ecological footprint that dynamically
 479 makes it possible to achieve statistical convergence. The lagged-dependent ecological footprint
 480 (EF_{t-1}) with a statistically significant positive autoregressive coefficient 0.752 is less than 1—
 481 implying a dynamically stable relationship between ecological footprint, income, squared of
 482 income, airline transport demand, rail transport demand, democracy, renewables, and fossil fuels.
 483 This infers that historical trends of ecological footprint predict 75% future changes in ecological

484 status, based on the *ceteris paribus* assumption. Thus, ecological destructive levels or patterns of
 485 ecological sustainability across countries are determined by historical consumption patterns of
 486 ecological resources. Consistent with the FMOLS estimator, an inversed-U-shaped relationship
 487 between ecological footprint and income level is observed. While a 1% increase in income level
 488 expands the ecological footprint by 15%, the squared component of income level declines
 489 ecological footprint by 0.71%. Increasing the share of fossil fuels in the total energy consumption
 490 escalates ecological footprint by 0.16%. However, diversification of the total energy consumption
 491 with a 1% share of renewables declines ecological footprint by 0.04%. Moreover, the summary of
 492 estimated findings from both the FMOLS estimator and panel bootstrap-corrected fixed-effects
 493 analysis is depicted in Figure 5.

494 **Table 7.** Heterogeneous dynamics of ecological footprint with bootstrap-corrected fixed-effects

	Coefs.	t	P>t	[95% Conf. Interval]	
EF_{t-1}	0.752*** (0.075)	10.050	0.000	0.604	0.900
Y	14.998*** (5.603)	2.680	0.008	3.910	26.086
Y^2	-0.712*** (0.263)	-2.710	0.008	-1.233	-0.191
AR	0.043 (0.027)	1.600	0.113	-0.010	0.097
RL	-0.006 (0.011)	-0.560	0.573	-0.029	0.016
D	0.015 (0.057)	0.270	0.790	-0.097	0.128
R	-0.044***	-3.230	0.002	-0.071	-0.017

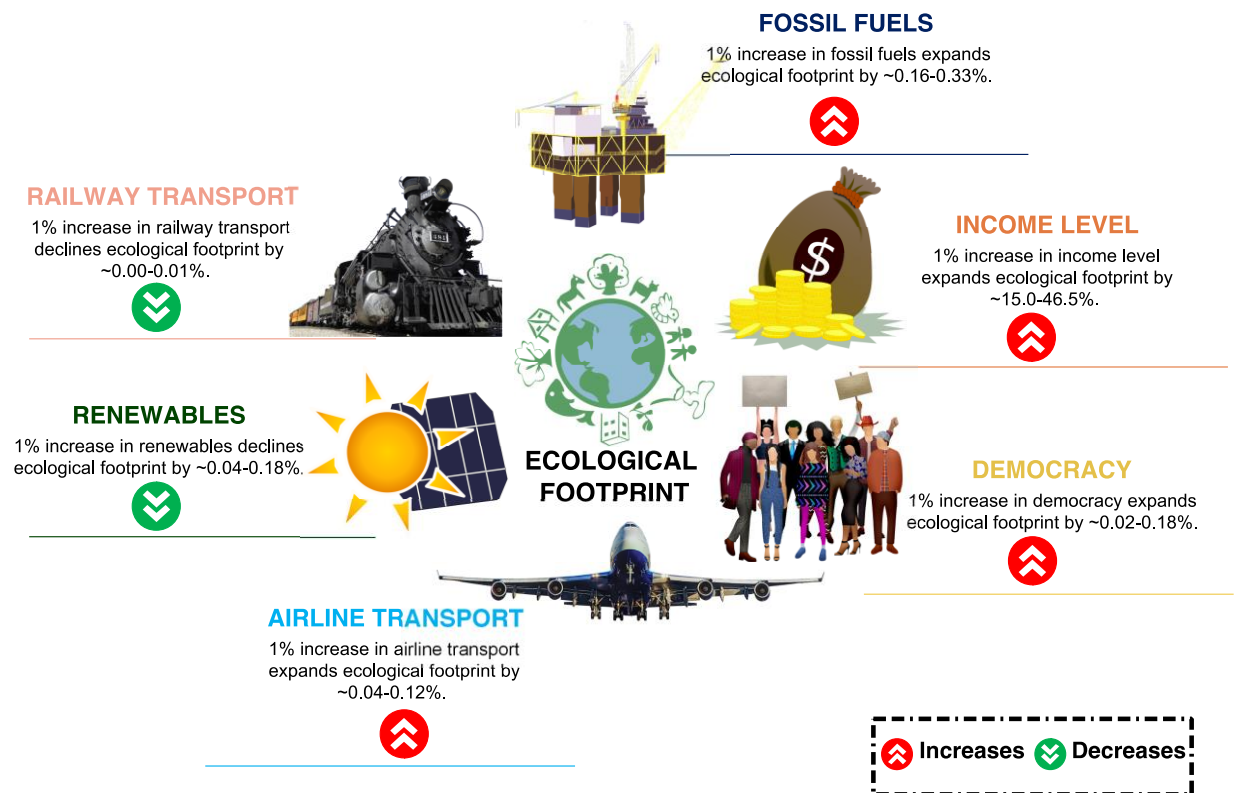
	(0.014)				
<i>F</i>	0.160**	2.410	0.017	0.029	0.292
	(0.066)				

495 **Note:** (..) denotes the bootstrapped standard errors, ***, and ** represent statistical significance at $p < 0.01$ and
496 $p < 0.05$. Confidence interval [95% Conf. Interval] for the t-distribution calculated with bootstrapped standard errors
497 whereas statistical inferences are performed with parametric bootstrap technique. **Legend:** *EF* represents ecological
498 footprint per capita, *Y* is the gross domestic product (GDP) per capita (constant, 2010 US\$), Y^2 is GDP per capita
499 square, *AR* is The *AR* is airline transport demand (number of passengers (carried)), *RL* is the railway transport
500 demand (number of passengers (carried)), *D* is the democracy index, *F* is the fossil fuel energy consumption (% of the
501 total), and *R* is renewable energy consumption (% of total final energy consumption)).

502 4. Discussion

503 This study executed the test of environmental sustainability among air travel demand, rail
504 transportation demand, demographic policies, and energy consumption from conventional energy
505 sources and renewable energy in G7. The nexus between these indicators has far-reaching
506 implications for environmental sustainability targets in G7 economies. For instance, the significant
507 positive interaction between income level and environmental degradation is suggestive that at
508 higher economic growth there is a compromise for environmental quality—as such, validating the
509 EKC phenomenon. Thus, this study gives credence to the scale effect phase of development in G7
510 over the investigated period, where the emphasis is on economic growth relative to environmental
511 status. This outcome is in line with the study of Balsalobre-Lorente (2018) for 5-EU countries. For
512 the case of G7 policymakers and energy specialists to pursue green-development strategies, an
513 economy that thrives on clean and sustainable energy that is conscious of the threshold of where
514 trade-off exists between economic growth becomes environmentally friendly (Bekun et al. 2019).
515 To maintain the momentum of environmental sustainability in the transportation sector, especially
516 those that stem from air and rail transport demand is affected by demographic policies amidst
517 global energy demand. There is a need for a sustainable transition to a full-fledged alternative and

518 clean energy sources (renewables, nuclear) and innovation in the transportation sector and
519 economic structure. For instance, there is the need for government officials and private investors
520 in the aviation sector to embrace low-carbon efficient planes that are in line with global renewable
521 expectations for a clean ecosystem—as already witnessed in some quarters of the world (Erdogan
522 et al., 2020). The need for more strides on renewable energy in G7 is evidence of the positive
523 relationship highlighted between fossil-fuel energy sources that deteriorate environmental quality.
524 As such, government administrators can pursue environmental sustainability by adopting the
525 polluter-pays principle—a phenomenon that emphasizes the need for regulation(s) on the violator
526 of environmental rules subject to cost damage. We further observe that government policies affect
527 environmental quality, as democratic-based policies exacerbate environmental quality. To achieve
528 the environmental sustainability target, there is a need for a revamp of government regulation(s),
529 as it concerns demographic indices highlighted in existing studies (Usman et al. 2019; Usman et al.
530 2020). Additionally, rail transport demand shows a desirable relationship with environmental
531 quality. To maintain the positive strides of renewable energy, it calls for responsible energy
532 consumption in G7 by diversifying the energy mix (SDG-11 and 12), which by extension aids to
533 climate change mitigation (SDG-13).



534

535 Figure 5. Summary of estimated findings from both FMOLS estimator and panel bootstrap-
 536 corrected fixed-effects analysis

537 5. Conclusion

538 Using annual frequency data for G7 economies, this study explored the connection between air
 539 travel demand, rail transportation demand, institutional policies, and energy consumption from
 540 conventional energy sources and renewable energy on environmental degradation measured by
 541 broader proxy, viz. ecological footprint. For sound and efficient analysis, panel techniques that
 542 account for cross-sectional dependence and heterogeneity were employed. A long-run equilibrium
 543 relationship was confirmed over the sampled period.

544

545 Our study's empirical results validated the existence of the EKC concept in G7, where there exists
 546 a trade-off between GDP growth and environmental quality. Our empirical results outline policy

547 implications for stakeholders—where there is the need for caution on green development for
548 environmental sustainability without compromising sustained economic development.
549 Additionally, we found fossil-fuel sources hamper environmental quality, thus, suggesting the need
550 for more innovative and clean energy sources like renewables (i.e., bioenergy, wind, photovoltaic,
551 hydro, solar, and thermal energy sources). There is also a need for drafting policies that engender
552 environmental sustainability in lieu of the democratic dynamics of G7 architecture. The revelation
553 of air transport-induced environmental pollution is indicative to aviation agents for alternative
554 innovation to conduct operations while reducing environmental degradation. This can be achieved
555 by the adoption of planes with energy-efficient, low-carbon emission, and clean technologies that
556 is more committed to environmental treaties and actions. On the contrary, rail transport and
557 renewable energy improve environmental quality and this momentum should be sustained for
558 environmental sustainability which is of utmost desire.

559

560 Conclusively, this study explored the nexus between air travel demand, rail transportation demand,
561 democracy, income level, and energy consumption from fossil fuels and renewable sources on
562 ecological footprint. However, subsequent studies can explore the theme and account for
563 asymmetry while considering other developed blocs. There is also limited documentation on
564 disaggregated data modeling which could be explored by future studies on the extension of the
565 EKC to the transportation sector-induced pollution.

566 **Declarations**

567 **Ethics approval and consent to participate:** Not applicable

568 **Consent for publication:** Not applicable

569 **Availability of data and materials:** The datasets used and/or analyzed during the current study
570 are available from the corresponding author on reasonable request.

571 **Competing interests:** The authors declare that they have no known competing financial
572 interests or personal relationships that could have appeared to influence the work reported in this
573 paper.

574 **Funding:** Not applicable

575 **Author Contributions:** **SE:** Formal Analysis, Writing - original draft; Writing - review & editing,,
576 **SAS:** Supervision, Writing - original draft; **FFA:** Writing-original draft, Writing - review &
577 editing. **FVB:** Conceptualization, Writing - original draft; Writing - review & editing, **PAO:**
578 Writing - original draft; Writing - review & editing,,

579

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